

Atmospheric deposition of mineral nitrogen and soil properties of forests in European Russia (Kostroma and Vologda regions)

I. Kudrevatykh, Russian Academy of Sciences, Russia
N. Ananyeva, Russian Academy of Sciences, Russia
K. Ivashchenko, Russian Academy of Sciences; Peoples' Friendship University of Russia, Agrarian Technological Institute, Russia

The Asian Conference on the Social Sciences 2016
Official Conference Proceedings

Abstract

In coniferous and mixed forests of Kostroma and Vologda regions of European Russia the sites (totally 39) were selected. In each site snow cores (March) and soil (August, soddy podzolic or Umbric Albeluvisols, 0-20 cm layer, litter excluded) samples were taken from a 10 m² plot by augering (corners and center). Bulk sample of snow and soil for each site was prepared. Ammonium (NH₄⁺) and nitrate (NO₃⁻) content in snow (melt water, atmospheric deposition) and soil was measured, mineral nitrogen (N_{min}) calculated as NH₄⁺+NO₃⁻. In the soil sample contents of total organic carbon (C_{tot}), total nitrogen (N_{tot}), P, Al, Ca; pH, texture and microbiological properties: microbial biomass carbon (C_{mic}) and microbial respiration (MR) were determined. The ratios of MR / C_{mic} = qCO₂ and C_{mic} / C_{tot} were calculated. Atmospheric deposition of N_{min} was reached on average 0.9 kg N ha⁻¹ yr⁻¹ and 4.5 kg N ha⁻¹ yr⁻¹ for Kostroma and Vologda regions, respectively. For studied regions a significant (p<0.05) positive correlation (r) between N_{min} and Al, Ca, MR, C_{mic}, C_{mic} / C_{tot} was found. Soil values (C_{mic}, MR, C_{mic} / C_{tot}, Al, Ca) in Vologda region were significantly (p<0.05) higher those in Kostroma region.

Keywords: atmospheric deposition, mineral nitrogen, soil, forest, microbial biomass, microbial respiration

iafor

The International Academic Forum
www.iafor.org

Introduction

During the last two centuries, agriculture and fossil fuels burning (anthropogenic activity) are resulted to the emission and atmospheric deposition of nitrogen compounds [nitric oxide (NO_x), nitrates (NO_3^-), ammonia (NH_3), ammonium (NH_4^+)]. Over the last 150 yrs. this deposition increased globally by 200% (Intergovernmental Panel..., 2001). Nitrogen compounds from atmosphere put into terrestrial ecosystems with wet (snow, rain) and dry (dust, mist, and etc.) depositions and estimated at present from 1 to 100 kg N ha⁻¹ yr⁻¹ (Galloway et al., 2008; Sutton et al., 2011; Jiang et al., 2012). Areas with high nitrogen deposition (≥ 20 kg N ha⁻¹ yr⁻¹) are subject to more attention of researchers in Western Europe, USA, Canada and China (Aherne, Posch, 2013; Sutton et al., 2014) and, in particular, within the Project “The Convention on Transboundary Air Pollution long distances”.

It was shown atmospheric nitrogen on average 2-3 times intensively compared to herbs and shrubs (Koptsik et al., 2008). Therefore, for forest ecosystems atmospheric nitrogen deposition might be a source of some environmental risk, which associated firstly with additional income of nitrogen to soil (MacDonald et al 2002; Manzoni et al, 2008). The consequence of such soil property changes could be an increase of primary plant production (de Vries et al., 2010; Bobbink et al., 2010) and respiration (mineralization) activity of soil microorganisms (Compton et al., 2004; Allison et al., 2008).

Nowadays atmospheric nitrogen deposition on European Russia is roughly, so we selected two areas with different deposition level. The aim of our study was focused on assessment of atmospheric nitrogen deposition and its influence on chemical and microbiological soil properties in forests of Kostroma and Vologda regions.

Materials and methods

Localization and sampling. The regions of Kostroma: 57°21'-59°36' N / 40°31'-47°05' E (few sources of atmospheric nitrogen: power stations, motor roads) and Vologda: 59°07'-61°02' N / 36°18'-39°57' E (plenty sources: production of ammonium fertilizer, timber, mechanical engineering, metallurgy, power stations, motor roads) were studied. Atmospheric deposition of nitrogen and some properties of soddy podzolic soil (Umbric Albeluvisols, WRB) were studied in deciduous, coniferous and mixed forests (40-60 yrs. age) of these two regions. Dominants of woody trees in Kostroma region were represented by *Picea obovata* Ledeb., *Abies sibirica* Ledeb., *Betula pubescens* Ehrh., in Vologda were *Calamagrostis arundinaceae*, *Trientalis europaea*, *Maianthemum bifolium*, *Pyrola rotundifolia*, *Dryopteris carthusiana*. The average annual air temperature is +3.0 and + 3.2°C, annual precipitation is 570 and 600 mm for Kostroma and Vologda regions, respectively.

The plots (10 m², 19 and 20 for Kostroma and Vologda regions, respectively) were selected at different distance from the main sources of atmospheric pollution (cities, factories, roads). In each plot the snow cores (center and corners) were collected by plastic sampler (diameter 5 cm, depth 50 cm) in March (the greatest snow accumulation). Bulk snow sample for each site was prepared, melted at 22°C and filtered through ash-free paper filter and fixed volume. It allow us to take into account

"wet" and "dry" atmospheric nitrogen deposition and thereby to estimate the accumulation of these pollutants for a long time.

Soil samples (upper mineral layer, 0-20 cm) were taken (center and corners) in August, bulk sample for each site was prepared (totally 39), delivered in the lab, air-dried (22°C), sieved (mesh 1 mm) and used further for chemical and microbiological analysis.

Methods. N_{MHH} content in melted water (December-January-February) was calculated using formula: $N'_{\text{min}} = N_{\text{min}} \times V$, where N'_{min} – nitrogen content in snow, kg N ha^{-1} ; N_{min} – nitrogen in melted water, mg N L^{-1} , V – melted water per one site (100 cm^2), L. In order to calculate atmospheric nitrogen deposition for whole year (snow, rain) the N'_{min} value should be multiplied by 4. Additionally in soil samples the contents of total carbon (C_{tot}), total nitrogen (N_{tot}) (analyzer Elementar Vario EL III) and elements of Al, Ca, P (x-ray fluorescence analysis, Spectroscan Max GV) were measured. Soil texture (pyrophosphate method) and soil pH (soil : water=1 : 2.5) were determined.

Soil microbial biomass carbon (C_{mic}) was measured by substrate-induced respiration (SIR) method, based on the maximal initial response (CO_2 production) on adding easily oxidized and available substrate (glucose, 10 mg g^{-1} , 0.1 mL g^{-1} soil) (Anderson and Domsch, 1978; Ananyeva, 2011). The C_{mic} ($\mu\text{g C g}^{-1}$) was calculated using the following formula: $C_{\text{mic}} = \text{SIR} (\mu\text{L CO}_2 \text{ g}^{-1} \text{ h}^{-1}) \times 40.04 + 0.37$ (Anderson and Domsch, 1978). Microbial respiration (MR) was measured in native soil samples (2 g, 24 h, 22°C, instead glucose the water added, 0.1 mL g^{-1} soil). The MR is expressed in $\mu\text{g CO}_2\text{-C g}^{-1} \text{ soil h}^{-1}$. The ratio of MR to C_{mic} (microbial metabolic quotient, $q\text{CO}_2$, $\mu\text{g CO}_2\text{-C mg}^{-1} C_{\text{mic}} \text{ h}^{-1}$) and $C_{\text{mic}} / C_{\text{total}}$ ratio (in %) were calculated. Preparation of soil samples prior microbiological analysis included moistening up to 55% of water holding capacity and pre-incubation (soil $\geq 150 \text{ g}$, 22°C, 7 d, air exchange).

Soil chemical and microbiological properties were performed in three replicates. The results were calculated for dry soil (105°C, 8 h) and expressed as the mean \pm standard deviation. Significance of difference ($p < 0.05$) of chemical and microbiological properties between two regions was tested by t-test. Relationship between atmospheric nitrogen deposition and soil properties was analyzed through Pearson's correlation coefficient. All experimental data were statistically processed and visualized (dot chart, box-plot, scatter plot) using R 3.2.4 software (<http://www.r-project.org/>)

Results and discussion

Atmospheric NH_4^+ deposition was ranged from 0.1 to 0.9 $\text{kg N ha}^{-1} \text{ yr}^{-1}$ in Kostroma region, it was higher (from 0.6 to 7.9 $\text{kg N ha}^{-1} \text{ yr}^{-1}$) in Vologda region (Fig. 1). However, atmospheric NO_3^- deposition in Kostroma and Vologda regions was not significantly ($p < 0.05$) differed (on average 0.6 and 0.7 $\text{kg N ha}^{-1} \text{ yr}^{-1}$, respectively). Besides, in Kostroma region atmospheric NO_3^- deposition was on average 2 times higher than that NH_4^+ , and opposite in Vologda region it was on average 5 times low. Atmospheric N_{min} deposition in Kostroma region was significantly ($p < 0.05$) low (on average 0.9 $\text{kg N ha}^{-1} \text{ yr}^{-1}$) compared to Vologda (on average 4.5 $\text{kg N ha}^{-1} \text{ yr}^{-1}$).

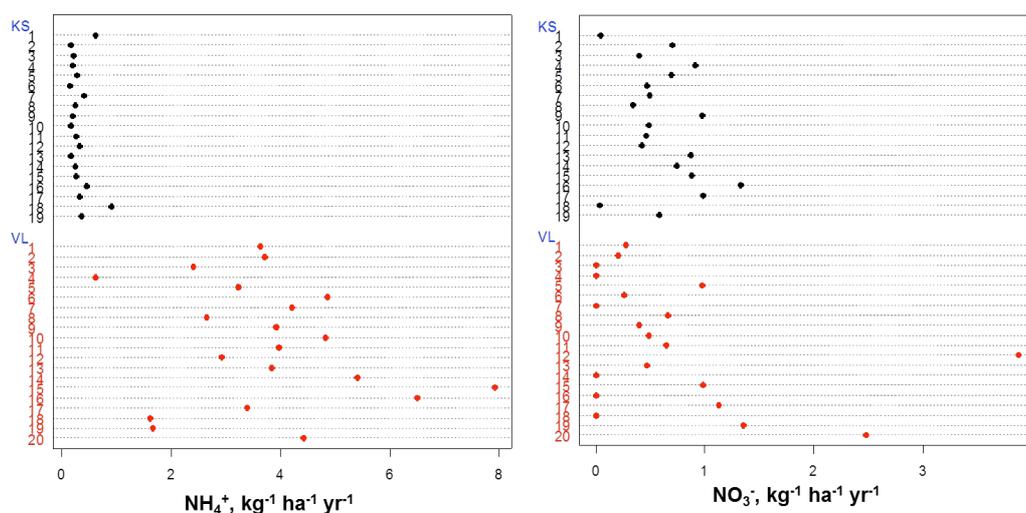


Figure 1. Dot chart of atmospheric ammonium (NH_4^+) and nitrate (NO_3^-) deposition for forest ecosystems of Kostroma (KS, 1-19 sites) and Vologda (VL, 1-20 sites) regions

Our previous research shown that atmospheric N_{min} deposition in forest ecosystems of Moscow region was varied from 4 to 11 $\text{kg N ha}^{-1} \text{yr}^{-1}$, on average 4.5 $\text{kg N ha}^{-1} \text{yr}^{-1}$ (Averkiewa, Ivashchenko, 2015). Atmospheric N_{min} deposition was ranged 5-10 and 10-18 $\text{kg N ha}^{-1} \text{yr}^{-1}$ for forests and industrial centers, respectively, of Croatia (1996-2008) (Alebic-Juretic, 2014), it was reached on average 0.6 and 9 $\text{kg N ha}^{-1} \text{yr}^{-1}$, respectively, in England and China (Cape et al., 2012; Jiang, 2013). Atmospheric N_{min} deposition for 200 observation areas (23 countries of Western Europe) was reached on average 14 $\text{kg N ha}^{-1} \text{yr}^{-1}$ (1.4-42 $\text{kg N ha}^{-1} \text{yr}^{-1}$), at that the authors consider the dependence of the fallout from urbanization (De Vries et al., 2010).

Soil properties (0-20 cm layer). Soil of studied regions was characterized by various texture (from light sandy loam to clay) and pH value, 3.3-6.5 units (Table 1). Content of soil C_{tot} , N_{tot} , NH_4^+ , NO_3^- , P and C / N, Al / Ca ratios in Kostroma region were not significantly ($p > 0.05$) differed compared to Vologda, however the contents of Al and Ca in Kostroma region were significantly ($p < 0.05$) less than those in Vologda (on average 1.2 and 1.4 times, respectively).

Table 1. Soil chemical properties (0-20 cm layer, range / mean) of Kostroma and Vologda regions (number of sites). Values with different letters significantly ($p < 0.05$) differ between regions (t-test)

Parameter	Kostroma (19)	Vologda (20)
N_{total} , %	0.04-0.43 / 0.20 a	0.05-0.56 / 0.24 a
C_{total} , %	0.60-6.30 / 2.98 a	0.93-10.47 / 3.73 a
C / N	13-19 / 15 a	12-26 / 16 a
$\text{pH}_{\text{вод}}$	3.3-6.5 / 5.1 a	3.5-5.6 / 5.2 a
Al, %	1.75-3.28 / 2.72 b	1.52-5.11 / 3.34 a
Ca, %	2.49-6.06 / 4.56 b	1.53-9.49 / 6.24 a
Al / Ca	5-10 / 6 a	3-12 / 6 a

	mg kg ⁻¹	
N-NH ₄	0.60-18.20 / 6.50 a	0.46-29.44 / 6.72 a
N-NO ₃	0-20.91 / 7.34 a	0.25-31.14 / 10.09 a
N _{MHH}	2.10-31.88 / 13.84 a	4.21-37.87 / 16.81 a
P	95-522 / 302 a	35-532 / 326 a

Soil C_{mic} and MR values of studied regions were varied from 51 to 1032 $\mu\text{g C g}^{-1}$ and from 0.2 to 4.3 $\mu\text{g CO}_2\text{-C g}^{-1} \text{ h}^{-1}$, respectively (Fig. 2). In soil of Kostroma region these values were on average 240 $\mu\text{g C g}^{-1}$ and 0.98 $\mu\text{g CO}_2\text{-C g}^{-1} \text{ h}^{-1}$, respectively, it was 2.4 and 2.0 times significantly ($p < 0.05$) less compared to Vologda region. The soil $q\text{CO}_2$ was on average 4.44 and 3.74 $\mu\text{g CO}_2\text{-C mg}^{-1} \text{ C}_{mic} \text{ h}^{-1}$ for Kostroma and Vologda regions, respectively. The C_{mic} / C_{total} ratio varied from 0.4 to 1.6% and from 0.9 to 2.5% in Kostroma and Vologda regions, respectively, it was on average (0.86 and 1.6%, respectively) significantly ($p < 0.05$) differed for Kostroma and Vologda regions.

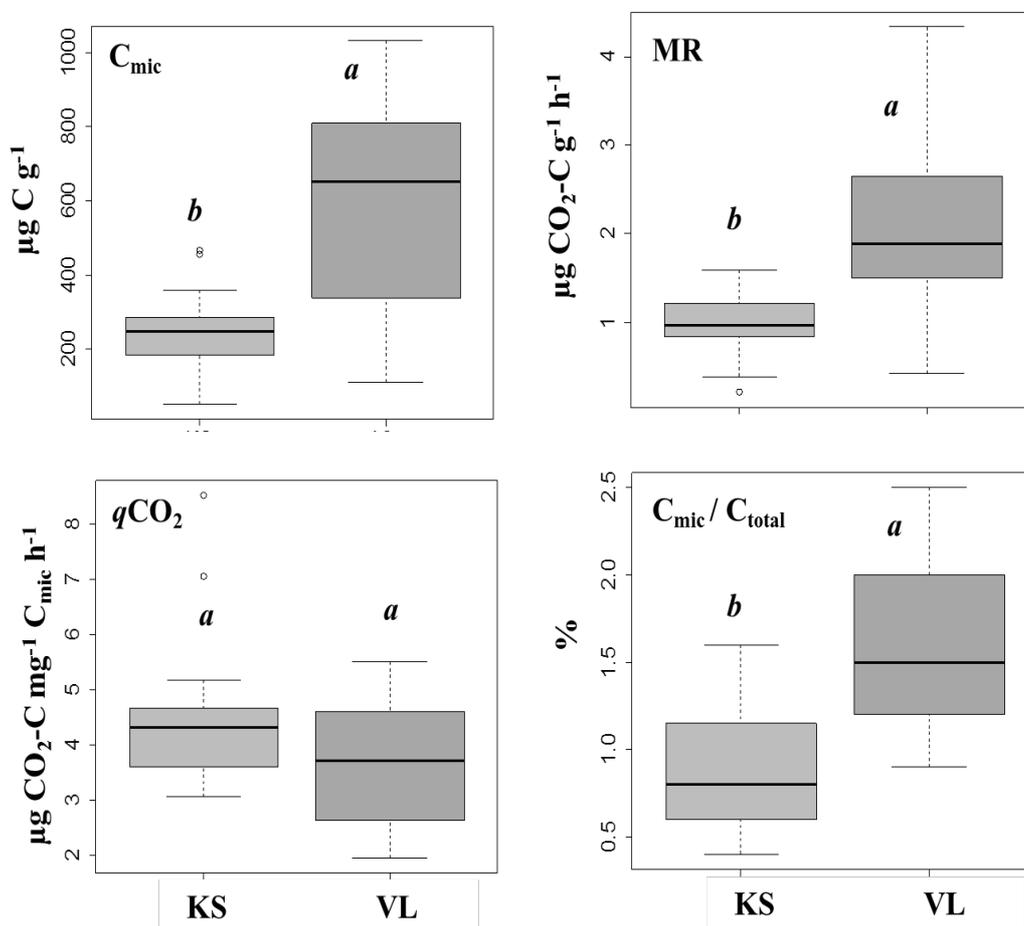


Figure 2. Soil microbial biomass carbon (C_{mic}), microbial respiration rate (MR), microbial metabolic quotient ($q\text{CO}_2$) and C_{mic} / C_{total} ratio in soils (0-20 cm) of Kostroma (KS) and Vologda (VL) regions. Values with different letters significantly ($p < 0.05$) differ between regions (t-test)

Box-plot: median; interquartile range, IQR (difference between the third, Q_3 , and the first, Q_1 , quartiles); lower whisker is $1.5 \text{ IQR} - Q_1$, upper whisker is $1.5 \text{ IQR} + Q_3$; outliers (circle) are values more and less upper and lower whiskers, respectively

Atmospheric N_{\min} deposition for two regions was significantly ($p < 0.05$), but not strongly, correlated with soil Al, Ca, C_{mic} , $C_{\text{mic}} / C_{\text{tot}}$ (Figs. 3, 4). Found that soil pH decreasing resulted atmospheric N_{\min} deposition might be caused the release and subsequent leaching Al and Ca from silicate minerals (Dise, 1995; MacDonald et al., 2002). Moreover, atmospheric N_{\min} deposition ($\leq 10 \text{ kg N ha}^{-1} \text{ yr}^{-1}$) was not be able to leach nitrates from soil, even in sandy soil (Neuvonen, 1990), and when it was $\geq 20 \text{ kg N ha}^{-1} \text{ yr}^{-1}$ this process was accelerated (Gundersen et al., 2006).

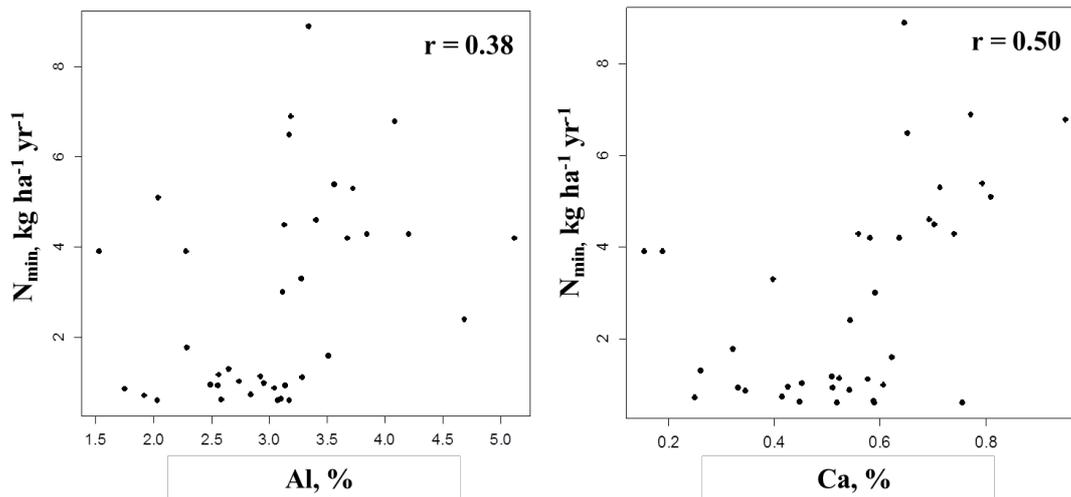


Figure 3. Relationship between atmospheric mineral nitrogen deposition (N_{\min}) and soil (0-20 cm) chemical properties (Al, Ca) of Kostroma and Vologda regions, 39 sites (r is Pearson correlation coefficient, $p < 0.05$)

An impact of atmospheric nitrogen deposition on soil microbial community functioning remains unclear, although additional nitrogen income in soil might be affected “directly” or “indirectly”. “Directly” impact of nitrogen might be considered as nutrition element for soil microbial biomass. “Indirectly” impact may be related to increasing forest primary production (Vitousek, Howarth, 1991; Magill et al., 2004) and changes of soil chemical properties: pH decreasing, C availability, Mg- or Ca-limitation and Al-toxicity (Compton et al., 2004; Treseder, 2008; Hu et al., 2010). Some authors shown that addition of N_{\min} to forest soils might be led to decrease soil microbial biomass and its respiration (Lee, Jose, 2003; Bowden et al., 2004; Compton et al., 2004; Frey et al., 2014), on the contrary, others authors found some increase of these parameters (Gallardo and Schlesinger, 1994). Moreover, there is an information that atmospheric N_{\min} deposition was not effected on soil microbiological properties (biomass, microbial respiration) of forest ecosystems (Nohrstedt, Börjesson, 1998; Brenner et al., 2005). We found the significant ($p < 0.05$) positive correlation ($r = 0.61$) between atmospheric N_{\min} deposition and soil $C_{\text{mic}} / C_{\text{tot}}$ ratio. It was found that increase of $C_{\text{mic}} / C_{\text{org}}$ or $C_{\text{mic}} / C_{\text{tot}}$ ratios (so-called index of C-availability) might be resulted to availability of soil C to microorganisms (Anderson, Domsch, 1986; Sparling, 1992). In our study the high atmospheric nitrogen deposition in forest ecosystems was increased soil C availability for soil microorganisms (increasing $C_{\text{mic}} / C_{\text{tot}}$ ratio).

Conclusion

Atmospheric N_{\min} deposition in Kostroma region was on average $0.9 \text{ kg N ha}^{-1} \text{ yr}^{-1}$, it was 5 times less in Vologda region (on average $4.5 \text{ kg N ha}^{-1} \text{ yr}^{-1}$). Umbric Albeluvisols C_{tot} , N_{tot} , $N\text{-NH}_4^+$, $N\text{-NO}_3^-$, N_{\min} , P contents were not significantly differed between two studied regions. Soil microbiological (C_{mic} , MR, $C_{\text{mic}} / C_{\text{tot}}$) and chemical (Al, Ca) properties of Umbric Albeluvisols in Kostroma region were significantly less compared to Vologda region. It might indicated, on the one hand, a more “favorable” functioning of soil microbial community and, on the other hand, some silicate minerals destruction of soil in Vologda region.

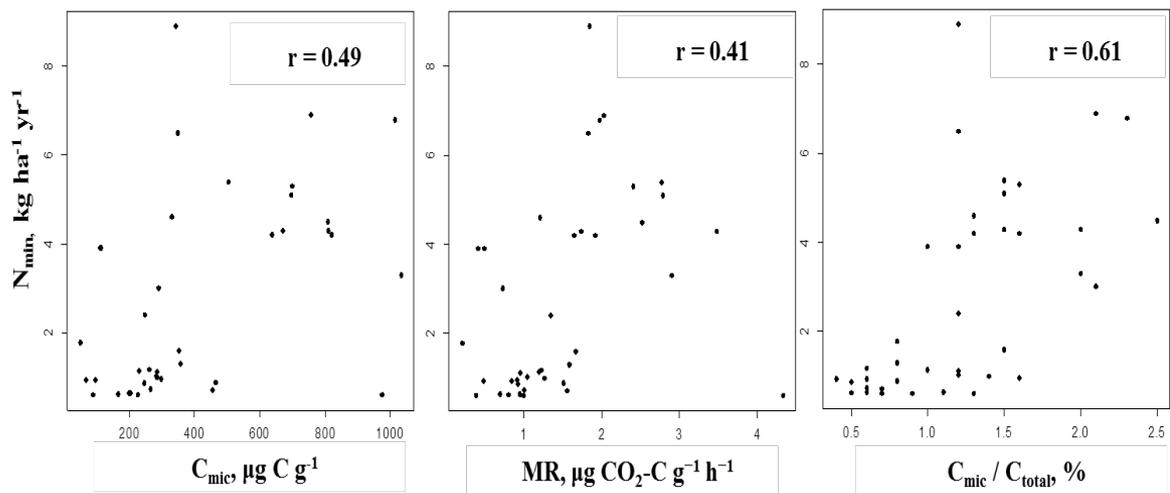


Figure 4. Relationship between atmospheric mineral nitrogen deposition (N_{\min}) and soil (0-20 cm) microbial biomass carbon (C_{mic}), microbial respiration (MR), $C_{\text{mic}} / C_{\text{total}}$ ratio of Kostroma and Vologda regions, 39 sites (r is Pearson correlation coefficient, $p < 0.05$)

This study was supported by Russian Foundation of Basic Research Grant No. 14-04-00098 and 15-04-00915.

References

- Aherne, J., Posch, M. (2013) Impacts of nitrogen and sulphur deposition on forest ecosystem services in Canada. *Current Opinion in Environmental Sustainability*, 5, 108–115.
- Alebic-Juretic, A. (2014) Nitrogen deposition within the littoral-highlands country of Croatia between 1996 and 2008. Nitrogen deposition, critical loads and biodiversity. M.A. Sutton, K.E. Mason, L.J. Sheppard, H. Sverdrup, R. Haeuber, W.K. Hicks. (pp. 67-73) London. Springer.
- Allison, S.D., Czimczik, C.I., Treseder K.K. (2008) Microbial activity and soil respiration under nitrogen addition in Alaskan boreal forest. *Global Change Biology*, 14, 1156–1168.
- Ananyeva, N.D., Susyan, E.A., Gavrilenko, E.G. (2011) Determination of the soil microbial biomass carbon using the method of substrate-induced respiration. *Eurasian Soil Science*, 44, 1215-1221.
- Anderson, T.H., Domsch, K.H. (1986) Carbon links between microbial biomass and soil organic matter. Perspectives in Microbial Ecology. (Eds): F. Megusar, M. Gantar. (pp. 467-471). Slovene Society for Microbiology. Ljubljana.
- Anderson, J.P.E., Domsch, K.H. (1978) A physiological method for the quantitative measurement of microbial biomass in soils. *Soil Biol. Biochem.* 10, 215-221.
- Averkjeva, I.Y., Ivashchenko, K.V. (2015) The Impact of anthropogenic emission of nitrogen on the functioning of forests in the European part of Russia. *Izvestiya Rossiiskoi Akademii Nauk. Seriya Geograficheskaya*, 2, 95-103.
- Bobbink, R., Hicks, K., Galloway, J., Spranger, T., Alkemade, R. et al. (2010) Global assessment of nitrogen deposition effects on terrestrial plant diversity: A synthesis. *Ecological Applications*, 20, 30–59.
- Bowden, R.D., Davidson, E., Savage, K., Arabia, C., Steudler, P. (2004) Chronic nitrogen additions reduce total soil respiration and microbial respiration in temperate forest soils at the Harvard Forest. *For. Ecol. Manage*, 196, 43-56.
- Brenner, R., Boone, R.D., Ruess, R.W. (2005) Nitrogen additions to pristine, high-latitude, forest ecosystems: consequences for soil nitrogen transformations and retention in mid and late succession. *Biogeochemistry*, 72, 257-282.
- Cape, J.N., Tang, Y.S., González-Benítez, J., Mitošinková, M., Makkonen, U., Jocher, M., Stolk, A. (2012) Organic nitrogen in precipitation across Europe. *Biogeosciences*, 9, 4401 – 4409.
- Compton, J.E., Watrud, L.S., Porteous, L.A., DeGroot, S. (2004) Response of soil microbial biomass and community composition to chronic nitrogen additions at Harvard forest. *Forest Ecology and Management*, 196, 143–158.

- Compton, J.E., Watrud, L.S., Porteous, L.A., DeGroot, S. (2004) Response of soil microbial biomass and community composition to chronic nitrogen additions at Harvard forest. *Forest Ecology and Management*, 196, 143-158.
- de Vries, W., Wamelink, G.W.W., van Dobben, H., Kros, J. et al. (2010) Use of dynamic soil-vegetation models to assess impacts of nitrogen deposition on plant species composition: an overview. *Ecological Applications*, 20 (1), .60 –79.
- Dise, N.B., Wright, R.F. (1995) Nitrogen leaching from European forests in relation to nitrogen deposition. *Forest Ecology and Management*, 71, 153 –161.
- Frey, S.D., Ollinger, S., Nadelhoffer, K., Bowden, R., Brzostek, E. et al. (2014) Chronic nitrogen additions suppress decomposition and sequester soil carbon in temperate forests. *Biogeochemistry*, 121, 305-316.
- Gallardo, A., Schlesinger, W.H. (1994) Factors limiting microbial biomass in the mineral soil and forest floor of a warm-temperate forest. *Soil Biology & Biochemistry*, 26, 1409-1415.
- Galloway, J.N., Townsend, A.R., Erisman, J.W., Bekunda, M., Cai, Z. et al. (2008) Transformation of the nitrogen cycle: Recent trends, questions and potential solution. *Science*, 320, 889 –892.
- Gundersen, P, Schmidt, I.K, Raulund-Rasmussen, K. (2006) Leaching of nitrate from temperate forests — effects of air pollution and forest management. *Environmental Review*, 14, 1-57.
- Hu, Y.-L. Zeng, D.-H., Liu, Y.-X., Zhang, Y.-L., Chen, Z.-H., Wang, Z.-Q. (2010) Responses of soil chemical and biological properties to nitrogen addition in a Dahurian larch plantation in Northeast China. *Plant Soil*. DOI 10.1007/s11104-010-0321-6.
- Intergovernmental Panel on Climate Change (2001) Climate Change 2001: The Scientific Basis. Contribution of Working Group I to the Third Assessment Report of the Intergovernmental Panel on Climate Change (eds Houghton JT, Ding Y et al.). Cambridge University Press, Cambridge.
- Jiang, C. M., Yua, W.T., Maa, Q., Xu, Y.G., Zou, H., Zhang, S.C., Sheng, W.P. (2013) Atmospheric organic nitrogen deposition: Analysis of nationwide data and a case study in Northeast China. *Environmental Pollution*, 182, 430–436.
- Koptsik, S., Koptsik, G., Alyabina, I. (2008) Assessment of the risk of excess sulfur input into terrestrial ecosystems of the Kola peninsula. *Russian Journal of Ecology*, 39, 5, 327–336.
- Lee, K.-H., Jose, S. (2003) Soil respiration, fine root production, and microbial biomass in cottonwood and loblolly pine plantations along a nitrogen fertilization gradient. *Forest Ecology and Management*, 185, 263-273.

MacDonald, J.A., Dise, N.B., Matzner, E., Armbruster, M., Gundersen, P., Forsius, M. (2002) Nitrogen input together with ecosystem nitrogen enrichment predict nitrate leaching from European forests. *Global Change Biology*, 8, 1028–1033.

Magill, A.H., Aber, J.D., Currie, W.S., Nadelhoffer, K.J., Martin, M.E. et al. (2004) Ecosystem response to 15 years of chronic nitrogen additions at the Harvard Forest LTER, Massachusetts, USA. *Ecological Management*, 196, 7-28.

Manzoni, S., Jackson, B., Trofymow, J.A., Porporato, A. (2008) The global stoichiometry of litter nitrogen mineralization. *Science*, 321, 684–686.

Neuvonen, S., Suomela, J. (1990) The effect of simulated acid rain on pine needle and birch leaf litter decomposition, *Journal of Applied Ecology*, 27, 857–872.

Nohrstedt, H.-Ö. & Börjesson, G. (1998) Respiration in a forest soil 27 years after fertilization with different doses of urea. *Silva Fennica*, 32(4), 383–388.

Sparling, G.P. (1992) Ratio of microbial biomass carbon to soil organic C as a sensitive indicator of changes in soil organic matter. *Austria Journal Soil Research*, 30. 195-207.

Sutton, M.A., Howard, C., Erisman, J.W., Billen, G., Bleeker, A. et al. (2011) The European Nitrogen Assessment. Cambridge: Camb. Univ. Press. 612 p.

Sutton, M.A., Mason, K.E., Sheppard, L.J., Sverdrup, H., Haeuber, R. et al. (2014) Nitrogen Deposition, Critical Loads and Biodiversity. Springer. 535 p.

Treseder, K.K. (2008) Nitrogen additions and microbial biomass: a meta-analysis of ecosystem studies. *Ecological Letter*, 11, 1111–1120.

Vitousek, P.M., Howarth, R.W. (1991) Nitrogen limitation on land and in the sea. How can it occur? *Biogeochemistry*, 13, 87-115.

Contact email: averkieva.irina@yandex.ru